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FINAL ECOLOGICAL RISK ASSESSMENT FOR AMERICAN CHEMICAL SERVICES GRIFFITH, INDIANA

MARCH 1992

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SECTION 7

SUMMARY

Based on this semi-quantitative, screening-level analysis of ecological risks, upland, wetland, and aquatic receptors may be adversely affected by contaminants present in the environmental media within the ACS watershed. The contaminants posing the greatest risk include PCBs and lead, which posed potential risks to all receptors examined. In addition, many of the metals, including mercury and zinc, as well as BEHP and heptachlor epoxide, posed potential risks to aquatic receptors and to mink.

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SECTION 6

RISK CHARACTERIZATION

The risk characterization integrates the exposure and toxicity assessments to estimate the potential hazard or risk to environmental receptors. Environmental media (e.g., sediment, surface water) concentrations or estimated daily intakes (EDIs) are compared with critical toxicity values (CTVs) by using a hazard index (HI).

Since media-specific criteria (e.g., tissue concentrations) were not available, hazard indices were calculated as follows:

HI = EDI / CTV

where:

EDI = Estimated daily intake of contaminant(s)

CTV = Critical toxicity value of the same contaminant(s)
through food, water, soil, sediment consumption.

Exposures through different environmental media were assumed to be cumulative. Consequently, the hazard index examines exposure to a contaminant through contact with all substantially contaminated media.

If an individual HI is greater than 1, it is suggested that the total exposure to a contaminant of concern through all exposure pathways is sufficient to produce a risk of adverse effects to the species of concern. However, if the individual HIs do not exceed 1, the risk may be negligible.

The following subsections discuss of the risks to aquatic life and terrestrial wildlife from contaminants at the ACS site.

6.1 RISK TO BURROWING RODENT

Risks to a burrowing rodent inhabiting the upland habitats found at the ACS site are presented in Table 6-1. Hazard indices exceeded unity for ethylbenzene, xylene, BEHP, PCBs, iron, lead, and zinc. For ethylbenzene and xylene, the primary potential exposure route

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TABLE 6-2 (Continued) Dose Estimation and Risk Charactization — Wetland Habitat Indicator Species — Mink

Chemicals of Potential Concern	Small Mammai Daily Intake (Di mammai)	Crayfish Daily Intake (DI cayfish)	Frog Daily Intake (DI frog)	Fish Daily intake (Di fish)	Water Daily Intake (Diwater)	Prey Delly Intake (Diprey)	Sediment Daily Intake (Disoli)	Total Daily Intake (Ditotal)	Critical Toxicity Value (CTV)	Hazard Index (HI)
	(mg/kg/day)	(mg/kg/day)	(mg/kg/day)	(m g/kg/day)	(mg/kg/day)	mg/kg/day)	(mg/kg/day)	(mg/kg/day)	(mg/kg/day)	(unitiess)
Volatiles .	i	•								
Benzene	2.98E-02	9.73E+01	3.89E+01	9.73E+01	2.70E+00	2.34E+02	1.14E-01	2.36E+02	6.8	`` 35
Toluene	0	0	0	0	0	0	1.38E-03	1.38E-03	46.6	2.98E-05
Ethylbenzene	0	~ 0	0	0	0	0	1.05E-03	1.05E-03	19.42	5.43E-05
Xylene	0	0	0	0	0	0	1.62E-03	1.62E-03	35.8	4.53E-05
2-Butanone	3.42E-03	8.74E+00	3.50E+00	8.74E+00	5.95E+00	2.10E+01	0	2.69E+01	92	3
Chlorobenzene	9.34E-05	3.31E-01	1.32E-01	3.31E-01	2.59E-03	7.94E-01	0	7.97E-01	5.45	0.15
Semivolatiles										
Bis(2-ethylhexyl)phthalate	1.85E-03	2.64E-01	7.64E+00	2.64E-01	1.35E-03	8.17E+00	3.57E-02	821E+00	0.38	. 22
Diethylphthalate	4.73E-06	1.61E-02	6.42E-03	1.61E-02	2.43E-04	3.85E-02	0	3.88E-02	150	2.59E-04
PAHe	0	0	0	0	0	0	7.71E-02	7.71E-02	7.14	0.01
PCBe	2.45E-02	1.12E+00	1.93E-02	2.86E+02	8.00E-04	2.87E+02	4.38E-02	2.87E+02	0.0015	191203
Heptachlor epoxide	0	0	0	0	0	0	5.35E-04	5.35E-04	0.00025	2
Inorganics										
Arsenio	0	0	0	0	0	0	1.29E-01	1 <i>2</i> 9E-01	024	0.54
Cadmium	2.01E-04	4.10E-02	1.64E-02	4.10E-02	8.38E-05	9.85E-02	3.81E-02	1.37E-01	0.1	1
Chromium (III)	0	0	0	0	0	0	2.00E+00	2.00E+00	293.6	0.007
Chromium (VI)	. 0	0	0	0	0	0	2 <i>2</i> 2E-02	2 <i>2</i> 2E-02	0.48	0.05
Copper	0	0	0	0	0	0	2.91E+00	2.91E+00	3.04	0.96
Iron	0	0	0	0	5.89E+00	. 0	0	5.89E+00	1.98	3
Lead	2.98E-04	3.32E-02	1.33E-02		124E-04	8.00E-02	5.69E+00	5.77E+00	0.05	115
Mercury	1.10E-04	1.86E+00	7.44E-01	1.86E+00	4.59E-05	4.47E+00	7.14E-02	4.54E+00	80.0	76
Zino	5.75E02	2.37E-01	8.26E+00		2.39E-02	2.92E+01	2.20E+00	3.14E+01	0.5	63
Cyanide	0	0	0	0	2.70E-04	0	0	2.70E-04	2.16	0.0001
TOTAL:										191525

DI sediment (mg/kg BW/day) = (CS: \times IR \times CF \times FI)/BW	EXPO:	SURE ASSUMPTION	ONS	
Di water (mg/kg BW/day) = (CW x CR)/BW	Water	Prey	Sedment	
DI prey (mg/kg BW/day) = (DI small mammal + DI crayfish + DI frog + DI fish)	0.025	150000	7500	ı
DI small mammal (mg/kg BW/day) = (Cew \times BCF \times 10E-3 \times % of diet \times CF \times FI)/ BW	1.0E-03	1.0E-08	1.DE-08	,
Di frog (mg/kg BW/dey) = (Cew \times BCF \times 10E-3 \times % of diet \times CF \times FI)/BW	1	1	1	
Di fish (mg/kg BW/day) = (Csw \times BCF \times 10E-3 \times % of diet \times CF \times FI)/ BW	0.925	0.925	0.925	
Di crayfish (mg/kg BW/day) = (Caw \times BCF \times 10E-3 \times % of diet \times CF \times FI)/BW				
Pecent of diet:				

IR = ingestion Rate (mg/day or L/day)
CF = Conversion factor (kg/mg)
FI = Fraction ingested (unitless)
BW = Body Weight (kg)

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Frog - 10% Fish - 25%

Small mammal - 40%

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was the consumption of forage grown in contaminated soil (maximum detected concentrations of these contaminants were found at sample location SB44-4.5). For iron, the primary exposure route was potentially through the ingestion of contaminated surface water from the drainage ditch (maximum detected concentration found at sample location SW05). For all other contaminants (lead, zinc, BEHP, and PCBs), the potential for exposure was from both consumption of browse grown in contaminated soil and incidental ingestion of contaminated soil. For lead, zinc, and BEHP, the maximum detected concentrations were found at sample location SB01-03, and for PCBs, the maximum detected concentration was found at sample location SB51-4.5.

6.2 RISK TO MINK

Risks to mink inhabiting the wetlands at the ACS site are presented in Table 6-2.

Hazard indices exceeded unity for benzene, 2-butanone, BEHP, PCBs, heptachlor epoxide, iron, lead, mercury, and zinc. For benzene, 2butanone, and BEHP, the primary exposure route is potentially through the ingestion of aquatic prey species that have bioconcentrated these contaminants from water (maximum detected concentrations at MW3-02, MW16-02, and MW17-01 for benzene, 2butanone, and BEHP, respectively). For iron, the primary exposure route is potentially through the ingestion of contaminated water (maximum detected concentration at MW16-02), and for heptachlor epoxide, the primary exposure route is potentially through the incidental ingestion of contaminated sediment (maximum detected concentration at sample location SD08). For all other contaminants (lead, mercury, zinc, and PCBs), the potential for exposure is through both ingestion of aquatic prey that bioconcentrated these contaminants from water and incidental ingestion of contaminated sediments. The maximum detected sediment concentrations were found at sample location SD16 for lead, mercury, and PCBs, and at SD14 The maximum detected concentrations in shallow groundwater were found at sample locations MW6-01, MW12-02, MW15-01, and MW17-01 for zinc, mercury, lead, and PCBs, respectively.

6.3 RISK TO AQUATIC RECEPTORS

Potential impacts on aquatic receptors can be evaluated by comparing contaminant concentrations in the drainage ditches and the shallow groundwater aquifer to ambient water quality criteria

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SECTION 5

TOXICITY ASSESSMENT

The toxicities of the contaminants of concern were assessed for aquatic life and for terrestrial life, as represented by the burrowing rodent and the mink. Based on a review of scientific literature, toxicity values were identified for the indicator or related species. These toxicity values are referred to as Critical Toxicity Values (CTVs). These CTVs range from the conservative No Observed Effect Level (NOEL) to the more drastic LD $_{50}$ (Lethal Dose to 50 percent of a test population). Criteria pertinent to the CTV selected for the species of concern represent the conservative end of this range.

In cases where CTVs were not available in the literature, they were derived from existing dose-response data. Critical Toxicity Values developed for laboratory animals were applied to assess the effects of site contaminants on the small mammal population in the open field and wetland habitats. To arrive at CTVs for the borrowing rodent and mink, available toxicity criteria were adjusted using conservative safety factors.

Conservative safety factors or uncertainty factors are applied to account for the uncertainty inherent in extrapolating available toxicity data. For those compounds for which only acute lethality values were available, toxicity values for this assessment were derived by dividing the acute toxicity value by appropriate safety factors. In evaluating the potential effects of pesticides on terrestrial species, U.S. EPA analyzed a subset of available dose-response data and suggested that if the estimated dose is less than one-fifth the median lethal dose for nonendangered species, no acute hazard can be presumed (Urban and Cook, 1986). This rule was adopted for this assessment and is presented in Table 5-1. A safety factor of 5 was used to account for extrapolating toxicity values for different species within the same class. In addition, a safety factor of 10 was used to adjust an acute lethality value to an acute no-observed effect level (NOEL), and to adjust a chronic or acute lowest observed effect level (LOEL) to a chronic NOEL (U.S. EPA, 1989). presents the CTVs for the burrowing rodent, and Table 5-3 presents the CTVs for the mink. If toxicity criteria were obtained from the literature, safety factors were not applied.

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Table 5-2 Critical Toxicity Values (CTVs) - Burrowing Rodent **American Chemical Services** Griffith, Indiana

	Toxic	city Data		Critical Toxicity	
Chemical of Potential Concern	Dose (mg/kgBW/day)	Effect/Species	Uncertainty Factor	Value (CTV) (mg/kgBW/day)	Source
Volatile Organics					
Benzene	3,400	LD ₅₀ /rat	500	6.8	Sax and Lewis, 1989
Toluene	233	NOEL/rat	5	146.6	IRIS, 1991
Ethylbenzene	97.1	NOEL/rat	5	19.42	IRIS, 1991
Xylenes	179	NOEL/rat	5	35.8	IRIS, 1991
2-Butanone	46	NOEL/rat	5	9.2	IRIS, 1991
Chlorobenzene	27.25	NOEL/dog	5	27.25	IRIS, 1991
Semivolatiles/Pesticides					
Bis(2-ethylhexyl)phthalate	19	LOEL/guinea pig	50	0.38	IRIS, 1991
PAHs (as naphthalene)	35.70	NOEL/rat	5	7.14	HEAST, 1991
PCBs	***	••		0.0015	Eisler, 1986b
Diethylphthalate	750	NOEL/rat	5	150	HEAST, 1991
Heptachior epoxide	0.0125	LOEL/dog	50	0.00025	IRIS, 1991
Inorganics					
Arsenic	1.2	NOEL/dog	5	0.24	ATSDR, 1991
Cadmium		-	-	0.1	Eisler, 1985

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Table 5-3

Critical Toxicity Values (CTV) - Mink American Chemical Services Griffith, Indiana

	Toxi	city Data		Critical Toxicity	
Chemical of Potential Concern	Dose (mg/kgBW/day)	Effect/Species	Uncertainty Factor	Value (CTV) (mg/kgBW/day)	Source
Volatile Organics					
Benzene	3,400	LD ₅₀ /rat	500	6.8	Sax and Lewis, 1989
Toluene	233	NOEL/rat	5	46.6	IRIS, 1991
Ethylbenzene	97.1	NOEL/rat	5	19.42	IRIS, 1991
Xylenes	179	NOEL/rat	5	35.8	IRIS, 1991
2-Butanone	46	NOEL/rat	5	9.2	IRIS, 1991
Chlorobenzene	27.25	NOEL/dog	5	5.45	IRIS, 1991
Semivolatiles/Pesticides					
Bis(2-ethylhexyl)phthalate	19	LOEL/guinea pig	50	0.38	IRIS, 1991
PAHs (as napthalene)	35.70	NOEL/rat	5	7.14	HEAST, 1991
PCBs	-	140	443	0.0015	Elsier, 1985
Diethylphthalate	750	NOEL/rat	5	150	HEAST, 1991
Heptachlor epoxide	0.0125	LOEL/dog	50	0.00025	IRIS, 1991
Inorganics					
Arsenic	1.2	NOEL/dog	5	0.24	ATSDR, 1991
Cadmium	-	-	-	0.1	Eisler, 1985

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Critical Toxicity Values expressed as body burden exposure (in mg/kg), were not available for fish. Rather, the toxicity of the contaminants in surface water to aquatic species can be assessed by comparing surface water concentrations in the drainage ditches and the wetlands to available acute and/or chronic AWQC or Lowest Reported Toxic Concentration values for the protection of freshwater aquatic life. AWQC are derived by U.S. EPA to protect 95 percent of all aquatic organisms, including fish, invertebrates, and aquatic plants. Lowest Reported Toxic Concentrations found in the literature were used for those chemicals for which the minimum data required to derive water quality criteria are not available.

Sediment quality guidelines can be used to assess the potential toxicity of sediments to benthic species. Various methods have been proposed to determine sediment concentrations associated with adverse biological effects, including the background approach, the sediment-water equilibrium partitioning (EP) approach, the spiked-sediment bioassay approach, the screening level concentrations approach, the Apparent Effects Threshold Approach, and the Bioeffects/Contaminant Co-Occurrence Analysis Approach. These approaches have been compared by the National Oceanic and Atmospheric Administration (NOAA) as part of the National Status and Trends program (NOAA, 1990).

The chemical concentrations observed or predicted by the different methods to be associated with biological effects were evaluated by NOAA, and the concentration at the low end (lower 10th percentile) of the range in which effects had been observed (ER-L) and the median data concentration in the range of reported values associated with biological effects (ER-M) were determined. These values were used to rank sites with regard to the potential for adverse biological effects. This range of data was for both marine and freshwater environments.

The Ontario Ministry of the Environment (Persaud et al., 1980) has established sediment quality guidelines for metals, nutrients, and organic compounds. These biological-effects-based guidelines can also be used as benchmarks to evaluate the potential impacts to benthic organisms. The guidelines define three levels of ecotoxic effects: no effect, lowest effect, and severe effect. A no-effect level is that level at which no effects have been observed in aquatic organisms. The lowest effect level is that level of sediment contamination that can be tolerated by the majority of benthic organisms. The severe effect level is that level at which

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Table 5-4

Sediment Concentrations Compared to Sediment Guidelines - Nonpolar Organics
American Chemical Services
Griffith, Indiana

				Sedi	ment Guid	elines (ppb))		
	Maximum Detected Concentration	EP-	*Safe	Level**	Effect	Range		Effects Ra	nge ^e
Contaminant	(ppb)	Approach ^b	Acute	Chronic	Low	Median	No	Lowest	Severe
Naphthalene	420	•						2,000	1.1E+07
2-Methylnaphthalene	380	_	-					2,000	1.1E+07
Pluorene	75	7,000	_					2,000	1.1E+07
Phenanthrene	440	1,390	14,000	_	_			2,000	1.1E+07
Anthracene	100	ı	•	_	_		_	2,000	1.1E+07
Fluoranthene	1,000	18,800	9,000	3,600			_	2,00 0	1.1E+07
Pyrene	1,100	13,100	49,500		-	-		2,000	1.1E+07 ~
Benz(a)anthracene	710	13,200	55,000	_	230	1,600	ı	2,000	1.1E+07
Chrysene	800	-	115,000	_	400	2,800		2,000	1.1E+07
Benzo(b)fluoranthene	1,500	-	4	_	-	1	-	2,000	1.1E+07
Benzo(k)fluoranthene	1,500	_	1	1	_	1	1	2,000	1.1E+07
Benzo(a)pyrene	600	10,630	450,000		400	2,500	_	2,000	1.1E+07
Indeno(1,2,3-cd)pyrene	420	_	1	_	_	-	_	2,000	1.1E+07
Dibenz(a,h)anthracene	200	_	240,000	_	60	260	_	2,000	1.1E+07
Benzo(g,h,i)perylene	550	_	_	_	_	_	_	2,000	1.1E+07
Bis(2-ethylhexyl)phthalate	4,400	-	1	3,240ª	_		_	_	1.1E+07
РСВ	5,400	_	35,900°	250°	50	400	10	70	530,000
Heptachlor epoxide	66	_	1.5ª	0.011ª	_		_	5	5,000

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Table 5-5 Sediment Concentrations Compared to Sediment Guidelines - Inorganics American Chemical Services Griffith, Indiana

			Sediment Gui	delines (mg/kg)	elines (mg/kg)		
Contaminant	Maximum Detected Concentration (mg/kg)	Effect Low	s Range ^a Median	Effect Lowest	Level ^b Severe		
			1				
Arsenic	15.9	33	85	6	33		
Cadmium	. 4.7	5	9	0.6	10		
Chromium, total	274	80	145	26	110		
Copper	359	70	390	16	110		
Lead	702	35	110	3	250		
Mercury	8.8	. 0.15	1.3	0.2	2		
Zinc	271	120	270	120	820		

Effects Range - The lower 10 percentile and median concentrations identified from chemical concentrations observed or predicted by different methods to be associated with biological effects (NOAA, 1990).

bEffects Level - Guidelines defining three levels of exotoxic effects; No effect level is that level at which no toxic effects have been observed on aquatic organisms; lowest effect level indicates a level of sediment contamination that can be tolerated by the majority of benthic organisms; severe effect level indicates the level at which pronounced disturbance of the sediment - dwelling community can be expected (Persaud, et al., 1990).

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For the mink, the following exposure assumptions were made:

- It was assumed that the home range of the mink is entirely within the wetlands.
- It was assumed that mink are exposed to contaminants via ingestion of surface water (upper aquifer concentrations), sediment (drainage ditches and wetlands), and prey from the wetlands.
- It was assumed that mink eat 40 percent small game, 25 percent fish, 25 percent crayfish, and 10 percent wetland amphibians.
- Contaminant concentrations in prey tissue were estimated as the product of the upper aquifer contaminant concentration and the BCF of the contaminant. As previously described, BCFs for the prey species of interest were obtained from the literature or derived from the log K_m of the contaminant.
- The following food and water ingestion rates and body weight, presented in the ACS Ecological Assessment (Warzyn, 1991), were applied:
 - Body weight = 925 grams (Burt and Grossenheider, 1980).
 - Water consumption rate = 25 mL/day.
 - Prey consumption rate = 150 grams/day.
 - Soil consumption rate = 7.50 g/day (assumes 5 percent of prey consumption rate).

4.3 AQUATIC RECEPTOR

Where toxicity data are expressed in terms of a medium concentration (e.g. Ambient Water Quality Criteria, Sediment Quality Guidelines), the determination of dose is not necessary. For the aquatic receptor, comparisons of predicted media concentrations (i.e., water and sediments) with media-specific toxicity data are made, since media-specific toxicity guidelines are available. These criteria and guidelines are levels that above which adverse effects to aquatic receptors have been observed.

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SECTION 4

EXPOSURE ESTIMATES

Tissue concentrations were not measured for the indicator species; rather, exposure doses were estimated at the point of contact using appropriate exposure algorithms. Exposure doses (in mg pollutant/kg body weight/day) were estimated from contaminant concentrations. Exposure concentrations used to calculate intake rates were the maximum detected contaminant concentrations in each medium that were presented in the RI (Warzyn, 1990). The maximum detected concentration was used in this risk assessment to represent the conservative case, at the request of the U.S. EPA BTAG. A conservative approach is followed in evaluating ecological risks at this site so that there is a high level of confidence in any no-impact conclusions made.

4.1 BURROWING RODENT

The estimated daily intake (EDI) of a contaminant by a burrowing rodent (Table 6-1) was estimated using the following algorithm:

 $EDI_{total} = EDI_{browse} + EDI_{soil} + EDI_{water}$

where:

EDI_{total} = Estimated total daily intake of contaminant (mg/kg BW/day)

EDI_{soil} = Estimated daily intake of contaminant via incidental soil ingestion

EDI_{unter} = Estimated daily intake of contaminant via water ingestion

EDI_{browse} = Estimated daily intake of contaminant via browse consumption

and:

EDI_{soil} = (Soil concentration x Soil intake rate) / Body Weight (BW)

EDI = (Water Concentration x Water intake rate) / BW

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EDI water Estimated daily intake of contaminant via water ingestion

Estimated daily intake of contaminant via incidental soil ingestion

EDI_{prey} Estimated daily intake of contaminant via prey consumption.

(Prey Concentration (C_{prev}) x % of diet) / BW

and:

(Water Concentration x BCF)

For aquatic organisms, bioconcentration is the process by which a compound is absorbed from water and concentrated by the organism to than the ambient concentration. In contrast, bioaccumulation is the process by which a compound is taken up by aquatic organism, both from water and through Bioconcentration factors (BCFs) for aquatic prey species were either obtained from the open literature or were derived from the octanol/water partition coefficient (K_{ow}) of the contaminant, using the following relationship from Vieth et al. (1980):

 $\log BCF = 0.76 \log K_{ou} - 0.23$

quantification Limited information on the of contaminant concentrations in edible terrestrial animal tissue in relation to environmental concentrations is available. For terrestrial prey species, BCFs were obtained from the literature or were derived using the following relationship from Kenega (1980):

 $log BCF = -3.825 + 0.701 log K_{m}$

For inorganic contaminants, it was assumed that 100 percent of the contaminant ingested from environmental media (e.g., water) was absorbed.

Bioconcentration factors for organic compounds derived from the above equations for aquatic and terrestrial organisms are presented in Table 4-1. The BCFs used in this exposure assessment for each species of interest are presented in Table 4-2.

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Table 4-2

Bioconcentration Factors (BCFs) by Species^a
American Chemical Services
Griffith, Indiana

Chemical of Potential	Macroin	vertebrates		Crayfish		Frog		Fish	<u>Small</u>	Mammals
Concern	BCF	Source	BCF	Source	BCF	Source	BCF	Source	BCF	Source
Organics										
Benzene	24	Vieth et al., 1980	24	Vieth et al., 1980	24	Vieth et al., 1980	24	Vieth et al., 1980	0.0046	Kenega, 1980
Toluene	70	Vieth et al., 1980	70	Vieth et al., 1980	70	Vieth et al., 1980	70	Vieth et al., 1980	0.012	Kenega, 1980
Ethylbenzene	146	Vieth et al., 1980	146	Vieth et al., 1980	146	Vieth et al., 1980	146	Vieth et al., 1980	0.024	Kenega, 1980
Xylene	138	Vieth et al., 1980	138	Vieth et al., 1980	138	Vieth et al., 1980	138	Vieth et al., 1980	0.023	Kenega, 1980
2-Butanone	0.98	Vieth et al., 1980	0.98	Vieth et al., 1980	0.98	Vieth et al., 1980	0.98	Vieth et al., 1980	0.00024	Kenega, 1980
Chlorobenzene	85	Vieth et al., 1980	85	Vieth et al., 1980	85	Vieth et al., 1980	85	Vieth et al., 1980	0.015	Kenega, 1980
Bis(2-ethyl- hexyl)phthalate	2,600 - 9,426	Verschueren, 1983	130	Verschueren, 1983	2,600-9,426	Verschueren, 1983	130	Verschueren, 1983	0.57	Kenega, 1980
Heptachlor epoxide	; 66	Vieth et al., 1980	66	Vieth et al., 1980	68	Vieth et al., 1980	66	Vieth et al., 1980	0.12	Kenega, 1980
PAHs (anthracene, BaP)	760	Verschueren, 1983	30-930	AQUIRE	760	Verschueren, 1983	30-930	AQUIRE	0.94	Kenega, 1980
PCBs	2,100-4,400	Verschueren, 1983	5.1 ^b	Charters, 1991	0.22 ^b	Charters, 1991	238,000	Verschueren, 1983	0.07 ^b	Charters

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SECTION 3

INDICATOR SPECIES

3.1 TERRESTRIAL SPECIES

Terrestrial habitats in the ACS watershed include approximately 1 to 2 acres of open field in the off-site containment area and the Kapica Drum property, approximately 33 acres of landfill open area, and 2 to 4 acres of wooded land along Colfax Avenue (Warzyn, 1990, Subsection 7.2.5.2). These areas are likely to support small herbivorous mammal populations, including various species of field rats, mice, voles, and woodchucks that live on the ground or burrow into or through it. A burrowing rodent was chosen as the indicator species in the open field habitat. A burrowing rodent may be exposed to site contaminants through direct ingestion of contaminated soil from the off-site containment area, consumption of surface water from the drainage ditches, and ingestion of forage grown in contaminated soils found in the off-site containment area.

The potential effects of site contaminants on wetland species was assessed by considering the mink (<u>Mustela vison</u>) as the indicator species. Although mink were not observed during the course of RI field activities, the U.S. F&WS and the U.S. EPA requested that this species be considered because of the potential mink habitat in the area and the availability of toxicological data for this species. In addition, the mink is a carnivorous wetland mammal that is sensitive to PCBs. A mink may be exposed to contaminants through the incidental ingestion of contaminated soil in the wetlands, consumption of shallow groundwater from the wetlands, and the consumption of contaminated prey inhabiting the wetlands.

3.2 AQUATIC SPECIES

The bluegill sunfish (<u>Lepomis macrochirus</u>) was selected as an appropriate aquatic indicator species because it is common in northern Indiana surface waters. The primary exposure routes of the bluegill include exposure to contaminants in surface water, sediments, and macroinvertebrates in the course of feeding.

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Table 2-2 (Continued)

Surface Soil and Sediment Concentrations American Chemical Services Griffith, Indiana

		Maximum Detected Concentration (mg/kg)	
Contaminant	Surface Soil (mg/kg)	Sediment (mg/kg)	Background Soil Concentration ^a (mg/kg)
Vanadium	1.1E+1	4.8E+1	1.9E+1
Zinc	2.9E+2	2.7E+2	2.8E+2
Cyanide, total	4.6E+0	<2.0E-3	3.7E+0

*From Table S-1 in ACS RI/FS (Warzyn, 1990).

NA - Not applicable; assumes background organic concentrations are zero.

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SECTION 2

CONTAMINANTS OF CONCERN

Chemicals of potential concern in each medium for each on-site habitat were selected based on a comparison of contaminant concentrations with toxicity criteria and background concentrations. persistence, as well as on the chemical bioaccumulation potential, and toxicity of the contaminant. Detailed information on the extent of contamination sampling at the ACS site is presented in Section 5 of the RI (Warzyn, 1990).

The on-site habitats and environmental media sampled in each were:

- Wetlands shallow aquifer groundwater, soil/sediments.
- Drainage ditches surface water, sediments.
- Terrestrial habitats off-site containment area surface soils.

Surface waters in the on-site drainage ditches were sampled as part of the RI (Warzyn, 1990, Section 5). Water samples were collected from the drainage from the off-site containment (SW05), the ditch west of ACS (SW07A), and the wetlands east of the landfill (SW08).

No surface water samples were collected directly from the wetlands. Since the wetlands receive groundwater discharge, contaminant concentrations in the shallow aquifer monitoring wells (MW-1 to MW-6 and MW-11 to MW-18) were used without dilution as estimates of the surface water concentrations in the wetlands.

The maximum detected surface water and shallow aquifer groundwater concentrations in both the drainage ditches and wetlands are compared with either Federal Ambient Water Quality Criteria (AWQC) for the protection of freshwater life or the Lowest Reported Toxic Concentrations found in the literature in Table 2-1. Lowest reported toxic concentrations are provided for chemicals for which the minimum data required to derive water quality criteria are not available. All contaminants that exceeded either value were considered to be contaminants of concern. Based on this zinc, cadmium, comparison, lead, iron, mercury, cyanide, polychlorinated biphenyls (PCBs), chlorobenzene,

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Table 2-1 (Continued)

Surface Water Concentrations American Chemical Services Griffith, Indiana

		Maximum Detected Concentration (µg/L)		er Quality a ^a (µg/L)
Contaminant	Shallow Aquifer- Wetlands	Drainage Ditches	Acute	Chronic
1,4-Dichlorobenzene	1.0E+1	<1.0E+1	1.12E+3	7.63E+2
1,2-Dichlorobenzene	· 3.3E+1	<1.0E+1	1.12E+3	7.63E+2
2-Methylphenol	3.8E+1	5.0E+0	None	None
Bis(2-chloroisopropyl)ether	3.0E+2	2.9E+1	2.38E+5	None
4-Methylphenol	2.2E+3	5.9E+2	None	None
Isophorone	3.5E+1	5.0E+0	1.17E+5	None
2,4-Dimethylphenol	1.1E+2	1.2E+1	2.12E+3	None
Naphthalene	7.1E+1	<1.0E+1	2.3E+3	6.2E+2
4-Chloro-3-methylphenol	5.0E+0	2.0E+0	3.0E+1	None
2-Methylnaphthalene	2.7E+1	<1.0E+1	2.3E+3	6.2E+2
Diethylphthalate	9.0E+0	<1.0E+1	9.4E+2	3.0E+0
Pentachlorophenol	3.0E+0	<5.0E+1	2.0E+1	1.3E+1
Di-n-butylphthalate	2.0E+0	<1.0E+1	9.4E+2	None
Bis(2-ethylhexyl)phthalate	5.0E+1	<1.0E+1	9.4E+2	3.0E+0
Benzoic acid	1.9E+3	8.5E+1	None	None
Pesticides/PCBs				
Aroclor 1248	2.6E+0	<5.0E-1	2.0E+0	1.4E-2
Aroclor 1260	2.7E+1	<1.0E+0	2.0E+0	1.4E-2
Inorganics				
Aluminum	2.8E+2	7.6E+2	None	None
Arsenic	4.32E+1	4.5E+1	3.6E+2	1.9E+2
Barium	1.84E+3	3.3E+2	None	None

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Table 2-1 (Continued)

Surface Water Concentrations American Chemical Services Griffith, Indiana

	Maximum Detected Concentration (µg/L)		Water Quality Criteria ^a (µg/L)		
Contaminant	Shallow Aquifer- Wetlands	Drainage Ditches	Acute	Chronic	
Zinc ^b	8.86E+2	8.8E+1	1.5E+2(2.2E+2)	1.4E+2(2.0E+2)	
Cyanide	1.0E+1	<1.0E+1	2.2E+1	5.2E+0	

⁴ Either Ambient Water Quality Criteria or Lowest Reported Toxic Concentration.

^c Assumes total chromium is 10% Cr (VI) and 90% Cr (III).

Sources: IRIS, 1991; Verschueren, 1983, U.S. EPA, 1989, U.S. EPA, 1986.

None - Criteria not available.

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hardness - dependent criteria: assumes 139 mg/L (Ca+Mg) for the drainage ditches and 210 mg/L (Ca+Mg) for the shallow aquifer; values in parentheses are criteria based on the shallow aquifer hardness.

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Table 2-2 Surface Soil and Sediment Concentrations American Chemical Services Griffith, Indiana

		m Detected tion (mg/kg)	
Contaminant	Surface Soil (mg/kg)	Sediment (mg/kg)	Background Soil Concentration ^a (mg/kg)
Volatile Organics			
Chloroethane	1.2E-2	4.0E-2	NA
Methylene chloride	1.9E-1	4.4E-2	NA
Acetone	8.7E+0	<1.0E-2	NA
Carbon disulfide	3.0E-3	<5.0E-3	NA NA
1,1-Dichloroethane	7.9E-1	<5.0E-3	NA .
1,2-Dichloroethene	2.6E+1	6.0E-3	NA .
Chloroform	3.0E-3	8.0E-3	NA NA
1,2-Dichloroethane	4.4E-2	<5.0E-3	NA
2-Butanone	9.0E+1	1.1E-2	NA .
1,1,1-Trichloroethane	5.6E-1	3.0E-3	NA .
1,2-Dichloropropane	3.5E-2	<5.0E-3	NA .
Trichloroethene ·	2.5E+2	<5.0E-3	NA
Benzene	2.3E+1	1.4E+1	NA NA
4-Methyl-2-pentanone	1.7E+1	<1.0E-2	NA NA
2-Hexanone	3.9E-1	<1.0E-3	NA .
Tetrachioroethene	2.4E+2	<5.0E-3	NA NA
Toluene	1.4E+3	1.7E-1	NA NA
Chlorobenzene	2.7E+1	<5.0E-3	NA NA
Ethylbenzene	5.7E+2	1.3E-1	NA
Styrene	2.6E+2	<5.0E-3	NA

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Table 2-2 (Continued)

Surface Soil and Sediment Concentrations American Chemical Services Griffith, Indiana

		m Detected tion (mg/kg)	
Contaminant	Surface Soil (mg/kg)	Sediment (mg/kg)	Background Soil Concentration ^a (mg/kg)
Butylbenzylphthalate	3.5E+0	1.7E-1	NA .
Benzo (a)anthracene	2.1E+0	7.1E-1	NA .
Chrysene	1.6E+0	8.0E-1	NA NA
Bis (2-ethylhexyl)phthalate	1.1E+2	4 4E+0	N A
Di-n-octylphthalate	3.8E+1	<3.3E-1	NA
Benzo (b) fluoranthene	3.5E+0	1.5E+0	NA NA
Benzo (k)fluoranthene	3.5E+0	1.5E+0	NA
Benzo(a)pyrene	1.4E+0	6.9E-1	NA
Indeno(1,2,3-c,d)pyrene	8.2E-1	4.2E-1	NA
Dibenz(a,h)anthracene	2.7E-1	2.0E-1	NA NA
Benzo(g,h,i)perylene	1.1E+0	5.5E-1	NA
Hexachiorobenzene	<3.3E-1	1.4E-1	NA
Pesticides/PCBs			
NAIdrin	8.8E-2	<8.0E-3	NA .
Endosulfan I	4.2E-2	<8.0E-3	NA NA
Heptachlor epoxide	<8.0E-3	6.6E-2	NA
Polychlorinated biphenyls	5.0E+1	5 4E+0	NA .

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endangered or threatened species. The King Rail, a state threatened species, was observed by the U.S. F&WS during a site visit. Other endangered or threatened species are suspected on the site based on observations of available habitat made by the U.S. F&WS.

The ACS site is not included as a designated area of special biological significance by the Indiana Department of Natural Resources (IDNR). Approximately 1.2 miles west of the site is the Hoosier Prairie State Nature Preserve, a relatively undeveloped property managed by the IDNR (Warzyn, 1991, Subsection 7.2.7.4).

Although the site is not designated as a special area in the Natural Heritage Program database, threatened or endangered species or unique plant communities could still exist on site. database is a growing database listing known sensitive areas. Important unknown areas are likely to exist in the state. reconnaissance survey by a trained biologist to determine whether sensitive species/communities are present has been recommended by U.S. EPA. Due to the season, this survey cannot be completed at this time.

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SECTION 1

INTRODUCTION

1.1 PURPOSE AND SCOPE

This ecological assessment is a screening-level evaluation of the environmental risks associated with contamination at the ACS site. This evaluation focuses on identifying potential adverse effects of contamination on flora and fauna in the nearby wetland and in onsite upland habitats. This assessment is based primarily on data presented in the ACS Remedial Investigation (RI) report prepared 1990 by Warzyn, Inc. in November (Warzyn, 1990) and the Final Draft Ecological Assessment prepared by Warzyn, Inc. in April 1991 (Warzyn, 1991).

This report is not intended to be a stand-alone document. This assessment uses the information presented in the ecological risk assessment prepared by Warzyn (Warzyn, 1991), while incorporating the comments made by the U.S. EPA Biological Technical Assistance Group (BTAG).

1.2 STUDY AREA DESCRIPTION

The ecological assessment of the ACS site examines an approximately 130 acre watershed in which the ACS site is located (Figure 7-3, Warzyn, 1991). The watershed lies between transportation corridors and consists of predominantly upland and wetland habitats. The RI indicates that this watershed is hydrologically isolated. Water sources are primarily rainfall, snowmelt, and groundwater discharge into the wetlands; discharge is primarily through evaporation and infiltration (Warzyn, 1991, Subsection 7.2.3.1)

1.2.1 Surface Water Features

Surface water features within the watershed include drainage ditches and industrial ponds. Surface water runoff is toward the west and south (Warzyn, 1990, Subsection 4.4.2).

A drainage ditch flows into the site at the northern boundary (directly north of the western ACS fence line), and then flows west along the northern site boundary and into the drainage ditch that cuts north to south through Wetland I. Another drainage ditch is

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